

**Trophic niche overlap and abundance reveal potential impact of interspecific interactions
on a reintroduced fish**

Sarah M. Larocque^{1*}, Timothy B. Johnson², Aaron T. Fisk¹

¹Great Lakes Institute for Environmental Research, University of Windsor, Windsor, Ontario,
N9B 3P4, Canada (SML: slarocque9@gmail.com; ATF: afisk@uwindsor.ca)

²Glenora Fisheries Station, Ontario Ministry of Natural Resources and Forestry, R.R. #4, Picton,
Ontario, K0K 2T0, Canada (TBJ: tim.johnson@ontario.ca)

* Corresponding author: Sarah M. Larocque

Ph: 1-519-253-3000 ext. 4931

Email: slarocque9@gmail.com

Abstract

Conceptually trophic niche overlap and species abundance can describe the strength and number of interspecific trophic interactions to determine the competitive impact on reintroduced species or other ecosystem changes. We use an example with young-of-year (YOY) Atlantic salmon (*Salmo salar*) reintroductions to determine if trophic niche overlaps and abundances limit restoration success. Using seasonal stable isotopes and abundance estimates for invertivorous fishes in three Lake Ontario tributaries, we assessed community isotopic structure, trophic niche overlap, and the impact of the niche overlap by incorporating relative abundance. Brown trout (*Salmo trutta*) YOY could be a strong competitor with a high trophic niche overlap with Atlantic salmon YOY but at lower abundances relative to Atlantic salmon minimizes impact. Stream resident fish communities appeared to partition resources across seasons such that abundant species had low trophic niche overlap to minimize overall competition with Atlantic salmon YOY given available resources, indicating niche complementarity. Through joint consideration of trophic overlap and abundance using our conceptual model, the competitive impact of community composition on a reintroduced species could be assessed.

Key words: stable isotopes; streams; salmonids; restoration; competition; Lake Ontario

Introduction

In aquatic systems, stocking fish to supplement wild populations and enhance fisheries occurs globally, and it is important to determine any limitations to stocking success (Minckley 1995; Brown and Day 2002; Molony et al. 2003). Removal of threats and having suitable habitat and food requirements are key factors to successful reintroductions (Harig et al. 2000; Cochran-Biederman et al. 2015). However, understanding a reintroduced species' interspecific interactions through potential competition can help determine biotic limitations to reintroduction success (Ewen and Armstrong 2007; Jachowski et al. 2016). The composition of the resident fish community may also influence the success of salmonid reintroductions (Harig et al. 2000; Scott et al. 2003; Cochran-Biederman et al. 2015). Not only can non-native species have negative influences on reintroduced species (Levin et al. 2002; Scott et al. 2003; Coghlan et al. 2007a) but so too can native species (Ward et al. 2008; Robinson and Ward 2011). Thus, existing resident species may be strong competitors with stocked species for food resources and/or habitat (Griffith et al. 1989; Schooley and Marsh 2007; Ward et al. 2008) and influence the outcome of the stocking.

The extent to which species overlap in trophic niches can indicate the degree of sharing of resources and potential interspecific competition (Hutchinson 1957), yet the intensity or impact of resource sharing and competition may in part be related to the abundance of the species involved, particularly in resource limited situations. Niche complementarity suggests that species co-exist by utilizing different resources (MacArthur and Levins 1967; Schoener 1974). However, few studies have assessed both niche partitioning and abundances simultaneously (e.g., Sugihara et al. 2003; Mason et al. 2008). When combining niche complementarity with abundance, there is evidence that abundant species would have low trophic niche overlap with

each other, thereby minimizing competition for resources and enhancing ecosystem function (Sugihara et al. 2003; Mason et al. 2008). However, we do not know whether this holds true during ecosystem disruptions such as the addition of new species and estimating the potential competitive impact on a species may be important for ecosystem management and understanding community structure.

We provide a conceptual framework to estimate the potential competitive impact on a species of interest by combining the strength of interaction based on the degree of trophic niche overlap with the number of interactions from the other species' abundance (Figure 1). In our conceptual framework, both trophic niche overlap and relative abundances are on a continuous scale from low to high, however, we considered the relative abundance of species in a community to be high if it is greater than the relative abundance of the species of interest. If trophic niche overlap among species is low, then the strength of the interspecific interactions are weak (Figure 1, bottom quadrants) and although the potential number of interactions with an abundant species can be high or frequent, it would be an example of niche complementarity and the resident species would have low impact on the introduced species (Figure 1 – bottom, right quadrant). The lower the relative abundance of the resident species with low trophic niche overlap the lower the impact on the introduced species (Figure 1 – bottom, left quadrant), while less abundant species with high trophic overlap would have a low impact due to the limited, but strong interactions with the introduced species (Figure 1 – top, left quadrant). However, if an introduced species has a high trophic niche overlap with an abundant species (Figure 1 – top, right quadrant), then it implies strong, frequent interactions with a species, and therefore potential competition for resources and a strong impact which could be a limitation on the

75 success of introduction. Incorporating the abundance of the species present in the system can
76 influence our interpretation of the importance of a high trophic niche overlap.

77 Trophic niche overlap is frequently assessed using stable isotopes as they provide better
78 estimates of trophic niches than stomach contents in describing longer term community or
79 population structure (Peterson and Fry 1987; Layman et al. 2007; Jackson et al. 2011). Isotopic
80 or trophic niche overlap has been previously used to assess diet partitioning between species and
81 potential indicator for competition (Wang et al. 2018; Ogloff et al. 2019). Similarly, metrics have
82 been used to describe the isotopic structure of the community (e.g., Layman metrics) and
83 whether it varies among communities with different interspecific diet partitioning and trophic
84 diversity (Layman et al. 2007; Sagouis et al. 2015). However, species abundance has yet to be
85 incorporated into conventional trophic niche evaluations which leaves a gap in understanding the
86 overall impact that the lack of diet partitioning may have on a species. Our conceptual
87 framework (Figure 1) provides a way in which the number of interactions based on other
88 species' abundances can be used in combination with the strength of isotopic trophic overlap to
89 determine the resulting impact to a species. This framework can be applied to various scenarios
90 in which changes in an ecosystem may warrant a better understanding of the trophic interactions
91 and resulting potential competitive impacts to a species by including abundance, such as
92 ecosystem disturbances or the addition of a species (e.g., native or invasive). Here, we use an
93 example with a reintroduced species to assess conventional isotopic analyses with niche overlap
94 and isotopic structure with resident species. We also incorporate species abundance with niche
95 overlap to better understand the overall impact of the community on the reintroduced species.

96 A species that may be negatively influenced through trophic interactions with resident
97 taxa is Atlantic salmon (*Salmo salar*) in Lake Ontario. Since Atlantic salmon extirpation in the

1890s, Pacific salmonids (*Oncorhynchus* spp.), among other species, have been introduced into Lake Ontario and elsewhere in the Laurentian Great Lakes, and different fish communities exist in the streams where Atlantic salmon young-of-year (YOY) are currently being reintroduced (Christie 1974; Crawford 2001). Studies have shown that non-native juvenile brown trout (*Salmo trutta*) and rainbow trout (*Oncorhynchus mykiss*; Stanfield and Jones 2003; Coghlan et al. 2007b) and native sculpin spp. (Gabler and Amundsen 2006) reduced growth and/or survival of Atlantic salmon YOY and can therefore be considered competitors with Atlantic salmon YOY and were a priori hypothesized to be potential competitors in Lake Ontario tributaries. Streams selected for Atlantic salmon reintroduction in the Lake Ontario watershed were chosen based on quality of suitable habitat (Stanfield and Jones 2003) but few studies have assessed the impacts of the local fish communities for potential competition for food resources (Johnson and McKenna 2015; Houde et al. 2016). Interspecific trophic competition may influence the growth, survival, and ultimately success of Atlantic salmon reintroductions in Lake Ontario. Thus, the stocking of Atlantic salmon YOY in Lake Ontario tributaries make for an ideal system to assess trophic interactions of a reintroduced species and the influence of the resident community.

Atlantic salmon YOY primarily consume aquatic macroinvertebrates (Keeley and Grant 1997; Coghlan et al. 2007a; Johnson 2013a, 2013b). By using stable isotopes to describe Atlantic salmon YOY diets, we can repeatably sample with non-lethal fin tissue (Sanderson et al. 2009; Hette-Tronquart et al. 2012) and determine how the trophic niche overlaps with the existing fish community of the same trophic guild. Dietary overlap with other salmonid and sculpin species may generate potential competition with Atlantic salmon YOY, and thus the degree of trophic niche overlap or trophic structure of different fish communities may influence Atlantic salmon reintroduction success. Diets within the fish community may also change seasonally depending

on food availability. Convergences in diet may occur during times in which food resources are more limited, such as overwinter as seen with YOY rainbow trout, brown trout, and Atlantic salmon (Johnson et al. 2017). Yet fish can also converge in diets during times with greater prey abundances and resource subsidies, such as terrestrial arthropods falling into streams in the summer (Nakano et al. 1999; Kawaguchi et al. 2001). Thus, seasonal variation in isotopic niche overlap may further influence the competitive impacts on Atlantic salmon.

In this study, we examine whether the potential success of a reintroduced fish species in streams may be limited by a strong overlap in food resources with other species (interspecific competition), in particular abundant species. The objective of this study was to determine the degree of food resource overlap and potential competition within three different invertivore fish communities seasonally by assessing species abundance, the communities isotopic trophic structure, and trophic niche overlap with Atlantic salmon YOY stocked into Lake Ontario tributaries. We hypothesized that 1) streams with more potential competitors (e.g., YOY brown trout, rainbow trout, and adult sculpin spp.) would have greater niche overlap with Atlantic salmon YOY and a smaller, condensed isotopic trophic structure, and 2) the greatest overlaps/smaller, condensed isotopic trophic structure would occur overwinter due to limited food resources and in summer due to abundance of terrestrial inputs. Lastly, we would determine whether Atlantic salmon were impacted (by having high niche overlap with abundant species) by resident fish species, in particular by potential competitors. These results will help determine the trophic interactions and relative impact with a reintroduced species over time and whether they may pose limitations to the successful reintroductions of Atlantic salmon in Lake Ontario.

Methods

Study area

The trophic niches of the invertivore fish community were assessed in three tributaries of Lake Ontario after being stocked with Atlantic salmon YOY by the Ontario Ministry of Natural Resources and Forestry (OMNRF). The study sites - Cobourg Brook (44.0315° N, 78.1453° W), Credit River (43.8024° N, 79.9959° W), and Duffins Creek (43.9483° N, 79.0802° W) – have similar cobble, gravel, and boulder dominated sediments and forested riparian vegetation, yet different flows, stream widths, as well as variation in fish communities at each site (Table S1). The OMNRF has stocked these sites in previous years and some Atlantic salmon that did not smolt in the spring may still be present as yearlings. All three sites have salmonids present, however, in terms of the presence of literature-based potential competitors (i.e., YOY brown trout, YOY rainbow trout, and adult sculpin spp.), Cobourg Brook had the most with brown trout YOY, rainbow trout YOY and adult slimy sculpin (*Cottus cognatus*), Credit River was intermediate with brown trout YOY and adult mottled sculpin (*Cottus bairdii*), and Duffins Creek had no potential competitors present.

Field Sampling

Between May 3 - 16 of 2017 Atlantic salmon YOY were stocked by OMNRF Normandale Fish Culture Station (Turkey Point, ON Canada) throughout the length of the three sites. Ten Atlantic salmon YOY from Normandale Fish Culture Station were euthanized and frozen for later stable isotope analyses to establish hatchery stable isotope values. Following stocking, at each field site, sampling occurred four times: 3 weeks post-stocking (spring – May/June 2017), summer (July/August 2017), fall (October 2017), and early spring the following year (overwinter – March 2018). Stream width (m) and water velocity (m/s) were measured at a representative location for each sampling event. Sampling consisted of using a 3-pass electrofishing removal method to estimate fish community abundances throughout the entire

167 stocking site. Using a backpack electrofisher (settings: 40 Hz, 250-450 V; make: Halltech
168 Aquatic Research, Guelph, ON, Canada) and two netters, each pass started at the downstream
169 end and worked upstream in a zigzag pattern, sampling all habitat types. All captured individuals
170 were processed and held in aerated coolers with stream water until all three passes were
171 complete prior to release. Fish processing involved identifying and counting all fish for each
172 pass, and a subset of fish per species ($n < 20$) were measured for fork length (± 1 mm) and mass
173 (± 0.1 g) and a small fin clip was taken from the upper caudal fin and stored frozen for later
174 stable isotope analyses. We used fin as a non-lethal alternative tissue to muscle for stable isotope
175 analyses to minimize impacting abundances through repetitive lethal sampling on small fishes
176 over the seasons (Sanderson et al. 2009; Hette-Tronquart et al. 2012). Small freshwater fishes
177 tend to have a strong linear relationship between fin and muscle, however, variation can exist
178 across species (Sanderson et al. 2009; Hette-Tronquart et al. 2012; Hayden et al. 2017). Only
179 potential invertivorous species with >5 individuals within a stream and occurring in at least two
180 seasons were further analyzed (with one exception, brook trout (*Salvelinus fontinalis*) YOY in
181 Duffins Creek that was only sampled in the fall). Species deemed to be primarily piscivores or
182 herbivores as determined in the literature (Scott and Crossman 1998) were not included for
183 further analyses.

184 *Stable Isotopes Analyses*

185 All fin clips were freeze-dried at -48°C for 48 h and cut to the appropriate size for
186 weighing (600 – 800 μg) into tin capsules. Stable isotope analyses were completed using a Delta
187 Plus isotope-ratio mass spectrometer (Thermo Finnigan, San Jose, CA, U.S.A.) coupled with an
188 elemental analyzer (Costech, Valencia, CA, U.S.A.). Standard delta notation (δ) was used to
189 express stable carbon ($\delta^{13}\text{C}$) and nitrogen ($\delta^{15}\text{N}$) isotope ratios in parts per thousand (‰)

differences from a standard reference material as the following equation: $\delta^{13}\text{C}$ or $\delta^{15}\text{N} =$
 $[(R_{\text{sample}}/R_{\text{standard}} - 1)] \times 1000$ where $R = {}^{13}\text{C}/{}^{12}\text{C}$ or ${}^{15}\text{N}/{}^{14}\text{N}$, respectively. Atmospheric nitrogen
 and Vienna Pee Dee Belemnite carbonate were the standard reference materials for ${}^{15}\text{N}$ and ${}^{13}\text{C}$,
 respectively. Precision was assessed by the standard deviation of replicate analyses of four
 standards (NIST1577c, internal lab standard (tilapia muscle), USGS 40 and Urea (n=125 for
 all)), which measured $\leq 0.22\text{‰}$ for $\delta^{15}\text{N}$ and $\leq 0.22\text{‰}$ for $\delta^{13}\text{C}$ for all the standards. Accuracy was
 based on the certified values of USGS 40 (n=125 for $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$) analyzed throughout runs
 and not used to normalize samples, which showed a difference of -0.01‰ for $\delta^{15}\text{N}$ and -0.02‰
 for $\delta^{13}\text{C}$ from the certified value. All stable isotope analyses were completed by the Chemical
 Tracers Laboratory at the University of Windsor's Great Lakes Institute for Environmental
 Research, ON, Canada.

Data were corrected for lipids since the elemental carbon-nitrogen (C:N) ratios in this
 study were above ≥ 3.4 for 94% of samples and ≥ 4 for 5% of samples. All samples with a C:N
 ratio > 3.4 , were lipid corrected using the Kiljunen et al. (2006) non-linear mathematical lipid
 normalization model with the Post et al. (2007) percent lipid calculation as recommended by
 Skinner et al. (2016).

Data Analyses

Age classes (YOY, yearling, adult) were assigned to salmon and trout based on length
 distribution curves in which distinct length-age classes were seen (generally YOY were < 125
 mm, yearlings were > 150 mm and < 250 mm, and adults were > 250 mm but this changed
 throughout the seasons). Brook trout was the only species with adults included in the analysis, as
 diet can include macroinvertebrates while other adult salmonids either had too few samples to be
 included (e.g., rainbow trout) or were considered primarily piscivorous which was confirmed

with elevated $\delta^{15}\text{N}$ (e.g., brown trout). Each lifestage for salmonid species were kept separated for analyses as YOY and yearlings can potentially be feeding on different items based on ontogenetic diet shifts (Keeley and Grant 1997; Mittelbach and Persson 1998).

To compare isotopic niche breadth between potential competitors (consumers of aquatic macroinvertebrates) and Atlantic salmon, we calculated the standard ellipse area for each species-lifestage in each stream and season using the SIBER (Stable Isotope Bayesian Ellipses in R) package in R (Jackson et al. 2011). We used the corrected version of the standard ellipse area (SEA_C), which contains ~40% of the isotopic data, thereby representing the core isotopic niche of each group while correcting for variable sample sizes (Jackson et al. 2011).

Using the maxLikOverlap function in SIBER, we estimated the maximum likelihood proportional isotopic niche overlap of Atlantic salmon YOY on species-lifestages (and species-lifestages niche overlap on Atlantic salmon YOY) within the same stream and season using species-lifestage-specific SEA_C . This provides two estimates of overlap depending on who is overlapping whom. For example, Atlantic salmon YOY could have 100% overlap with blacknose dace (*Rhinichthys atratulus*) whereas blacknose dace may only have 22% overlap with Atlantic salmon YOY depending on the position and size of each species-lifestage SEA_C . We included both overlap estimates to assess for trends. Spring was excluded from the seasons when comparing overlaps with Atlantic salmon YOY as the isotopic signature had not equilibrated to the stream diet at this time (i.e., isotopes reflected hatchery feed; see Supplementary Materials). This was to be expected as the isotopic value of fin can take approximately 2-3 months to equilibrate to their diet (or 4-5 tissue turnover half-lives) for rapidly growing, juvenile fishes (McIntyre and Flecker 2006; Heady and Moore, 2013). The $\log(x+1)$ -transformed proportion Atlantic salmon YOY overlaps onto species-lifestages (and vice versa) were compared among

streams and season (excluding spring) with a two-way ANOVA without an interaction term as we were only interested in the main effects.

For each stream by season, the relative abundance of each species-lifestage was calculated by the catch per unit effort (CPUE; number captured per electrofishing seconds) divided by the total fish CPUE over the sampling event. To assess for any trends in the catches, both the total fish CPUE and the relative abundance of Atlantic salmon YOY were compared across streams and seasons using a two-way ANOVA (without an interaction). The relative abundance of each species-lifestage was then plotted by the trophic niche overlaps with Atlantic salmon YOY and each species-lifestage. Proportional trophic niche overlaps greater than 0.5 were considered high. Any species-lifestage relative abundances that were greater than the minimum relative abundance of Atlantic salmon YOY from all sampling events (0.12) was considered high. The relative competitive impact of each species-lifestage on Atlantic salmon YOY was estimated depending on where the species-lifestage falls relative to the conceptualized framework quadrants.

Using the species-lifestages mean $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ values, community-level niche space metrics (also known as Layman's metrics) were calculated for each stream by season following Layman et al. (2007) and adopting a Bayesian approach implemented by Jackson et al. (2011) using the laymanMetrics function in SIBER. The community (stream by season) was considered significantly different for a metric when the 95% Bayesian credible intervals did not overlap another stream by season. The following metrics were used: the total area of the convex hull, the mean distance to the centroid, the range of carbon and nitrogen ($\delta^{13}\text{C}$ range and $\delta^{15}\text{N}$ range, respectively), the mean nearest neighbour distance, and the standard deviation (SD) of the nearest neighbour distance. An in-depth description for each metric can be found in Layman et

al. (2007). These metrics were used to indicate the total extent of carbon and nitrogen resources exploited by the community and the distribution of species-lifestages within the isotopic space.

All analyses were conducted in R version 3.6.1 (R Core Development Team, 2019) and significance was assessed at $\alpha = 0.05$ or whether 95% credible intervals did not overlap. Values are reported in mean \pm SD. Assumptions of normality and homoscedasticity were visually assessed using qqplot and fitted versus residual plots.

Results

A total of 688 samples were analyzed for stable isotopes from the three tributaries collected between May 2017 to March 2018 over four sampling events. Atlantic salmon YOY sizes ranged from a minimum mean fork length of 55 ± 5 mm (Duffins Creek, May 2017) to a maximum mean fork length 117 ± 10 mm (Credit River, March 2018; Table S2). Atlantic salmon YOY were generally the most abundant fish in each stream for every season, followed by either slimy sculpin (Cobourg Brook), mottled sculpin (Credit River), or longnose dace (*Rhinichthys cataractae*; Duffins Creek; Table S2). The total CPUE did not vary by stream ($F_{2,6} = 5.139$, $P = 0.050$) but differed across seasons ($F_{3,6} = 19.133$, $P = 0.002$) in which total CPUE was lower overwinter ($P < 0.01$) than any other season ($P > 0.05$; Figure 2; Table S2). Further, the relative abundance of Atlantic salmon YOY did not vary by stream or season ($F_{2,6} = 3.595$, $P = 0.0941$; $F_{3,6} = 3.856$, $P = 0.0751$, respectively; Figure 2). The standard ellipse areas (SEAc) for each species-lifestage by stream and season showed a distinct separation of the initial spring Atlantic salmon YOY from all other fish from the same stream community and was fairly similar to hatchery conspecifics (Figure 3). However, by summer (fall for Cobourg Brook), Atlantic salmon YOY ellipses had isotopic signatures aligned with the stream community (Figure 3). Otherwise, no distinct patterns were evident (Figure 3).

282 The proportion of Atlantic salmon YOY isotopic niche overlapping with other species-
283 lifestages niches within the same stream and season (excluding spring) ranged from 0 to 1 (Table
284 S2). However, Atlantic salmon YOY overlaps with species-lifestages was not significantly
285 different among streams ($F_{2,40} = 0.891$, $P = 0.418$) or seasons ($F_{2,40} = 0.724$, $P = 0.491$; Figure 4).
286 The isotopic niche overlap of other species-lifestages onto Atlantic salmon YOY also ranged
287 from 0 to 1 (Table S2). Similarly, species-lifestages overlapping with Atlantic salmon was not
288 significantly different among streams ($F_{2,40} = 0.730$, $P = 0.488$) or seasons ($F_{2,40} = 0.721$, $P =$
289 0.492 ; Figure 4). Across all co-occurrences with Atlantic salmon YOY (with either Atlantic
290 salmon YOY overlapping species-lifestages or vice versa), 28% ($n = 13$ of 45) had a high niche
291 overlap (proportion > 0.5) with Atlantic salmon YOY, which primarily occurred in Credit River
292 ($n = 6$) and Cobourg Brook ($n = 5$; Table 1). Atlantic salmon yearlings, blacknose dace, brown
293 trout YOY and yearlings, rainbow trout YOY, and rainbow darter (*Etheostoma caeruleum*) all
294 had a high niche overlap with Atlantic salmon YOY in at least one co-occurrence, and brown
295 trout YOY had the greatest proportion of occurrences with a high niche overlap with Atlantic
296 salmon YOY (Table 1). Brook trout adults and YOY, longnose dace, mottled and slimy sculpin,
297 and white sucker (*Catostomus commersonii*) had low or no niche overlap with Atlantic salmon
298 YOY (Table 1; Table S2).

299 When comparing the proportion of trophic niche overlap (Atlantic salmon YOY
300 overlapping species-lifestages, and vice versa) to relative abundance of the species-lifestages,
301 there was an absence of species-lifestages that had a high overlap with Atlantic salmon and high
302 relative abundance (proportion > 0.12), aside from rainbow darter from Duffins Creek in which
303 the relative abundance of Atlantic salmon in the same season (fall) was 0.4 (Figure 4; Figure 5).
304 The majority of species-lifestages had low niche overlap with and low relative abundance to

Atlantic salmon YOY with a few instances with high niche overlap and low relative abundance or low niche overlap and high relative abundance (Figure 5).

When assessing the Bayesian Layman metrics among streams and seasons, there was a relatively consistent pattern difference in which spring for all three streams had higher $\delta^{13}\text{C}$ range and mean distance to centroid than the other seasons within the same stream. The mean nearest neighbour distance and SD of the nearest neighbour distance also showed the similar pattern with higher values in spring than the other seasons but it was not significantly different for Duffins Creek (Figure 6). The total convex hull area had high values in spring but it was not significantly different from at least one other season within the same stream. The $\delta^{15}\text{N}$ range did not show any apparent distinction between seasons or streams (Figure 6). There were no significant differences between stream communities for any of the Layman metrics.

Discussion

We assessed the trophic niches of stocked Atlantic salmon YOY in multiple streams over seasons to determine whether potential competition with resident fishes may be occurring in Lake Ontario tributaries. Streams with greater or fewer a priori hypothesized potential competitors such as brown trout, overall did not appear to influence the isotopic niche space of Atlantic salmon YOY in any season. Using conventional stable isotope analyses, brown trout YOY strongly overlapped with Atlantic salmon YOY, however, with the inclusion of their low abundance it would not be perceived to have a large impact on Atlantic salmon YOY population growth or survival, as seen in our conceptual model. Combining abundance with stable isotope analyses revealed the impact of trophic interactions on a reintroduced species and provided insights on potential competition. Stream fish communities partitioned available resources such that trophic interactions were reduced with the stocked Atlantic salmon YOY, as evidenced by

the low isotopic niche overlap with abundant species. The low abundance of species-lifestages with a high niche overlap with Atlantic salmon YOY also indicates that there would be a low competitive impact to Atlantic salmon YOY populations. Further, there were no seasonal patterns (from summer to overwinter) in the relative overlap or niche size that would influence trophic interactions nor any changes in the relative abundance of Atlantic salmon YOY within the community. However, overwinter had the lowest total CPUE across streams due to low temperatures/conductivity which made electrofishing less effective. Across streams, the spring season had increased trophic niche breadth in the isotopic structure via Layman metrics, however, that was related to the isotope signatures of the stocked Atlantic salmon YOY not yet reaching dietary equilibrium with the stream from previous hatchery food. Our study suggests potential dietary competition from the invertivorous fish communities should not impact juvenile Atlantic salmon populations in these streams, particularly if juvenile Atlantic salmon are stocked in high abundances.

Isotopic niche overlap was not greater and isotopic structure was not smaller/condensed overwinter or summer relative to other seasons as hypothesized. Temporal reductions in food availability, like in winter, can increase niche overlap between species of the same trophic guild (Shustov et al. 2014; Sánchez-Hernández et al. 2016). Similarly, periods of high food availability, like in summer, can increase trophic niche overlap when species feed on similar abundant resources (Kawaguchi et al. 2001; Chen et al. 2011). With the abundant stocking of Atlantic salmon in streams, resources were likely limited during all seasons as abundant species-lifestages had distinct trophic niches and partitioned resources. Resource limitation in all seasons may explain the lack of seasonal differences in niche overlaps and isotopic structure.. Macroinvertebrate communities also change seasonally (Giller and Twomey 1993; Fierro et al.

2015), which will influence prey availability if fish species have certain prey preferences, and therefore degree of overlap with other fish may also be seasonally variable. As isotopes reflect a longer-term diet relative to stomach contents, as was seen during the post-stocking spring when Atlantic salmon YOY isotopes had yet to reflect stream isotopes, this method may not have been the most effective to capture shorter-term within season changes in diet. Instead investigating the stomach contents of invertivorous fish with the abundance and seasonality of macroinvertebrate can confirm resource limitation and further understand the seasonality of trophic niche overlaps with Atlantic salmon YOY.

Of the few species with high overlap with Atlantic salmon YOY, brown trout YOY had a high niche overlap with Atlantic salmon (and vice versa) for most of their co-occurrences. There may be competition for food resources with Atlantic salmon and brown trout YOY in Lake Ontario tributaries, however, it did not always occur in all seasons for some streams (e.g., Credit River – fall). Brown trout had more co-occurrences of high overlap with Atlantic salmon YOY than rainbow trout, potentially indicating that rainbow trout is not as frequent a competitor with Atlantic salmon YOY. Neither sculpin species had a high overlap with Atlantic salmon YOY suggesting they are not trophic competitors, as has been seen in other systems (e.g., Gabler and Amundsen 2006). Across multiple studies, when co-occurring, brown trout consistently reduced Atlantic salmon growth in artificial streams, while rainbow trout had variable results (Van Zwol et al. 2012; Houde et al. 2015a, b, 2017). Brown trout YOY have been shown to strongly compete with Atlantic salmon YOY for habitat resources (Armstrong et al. 2003) and feed on similar aquatic invertebrates (Dineen et al. 2007; Johnson and McKenna 2015). However, brown trout may feed more heavily on terrestrial invertebrates in Lake Ontario tributaries (Johnson and McKenna 2015), which may explain the one instance in which brown trout YOY did not highly

374 overlap with Atlantic salmon and they fed on different seasonal prey. Our study indicates that
375 competitive interactions between Atlantic salmon and brown trout likely exist, yet unlike other
376 studies with equal abundances of both species (e.g., Van Zwol et al. 2012; Houde et al. 2015a, b,
377 2017), our systems had low relative abundance of brown trout YOY and high relative abundance
378 of Atlantic salmon YOY such that the overall impact of any potential competition with Atlantic
379 salmon YOY would be low.

380 Together, trophic niche overlap and abundance can reveal instances where resource
381 sharing could infer competition and the resulting level of impact on a species. We used the
382 example with Atlantic salmon YOY to show how combining trophic niche overlap and
383 abundance indicated that there was no high impact towards reintroduction success. Abundance
384 has not been utilized in conjunction with stable isotopes, however, by doing so one can infer the
385 potential level of impact through competition for food resources that conventional analyses may
386 not reveal. In all three streams, abundant species did not have a strong niche overlap with the
387 abundant Atlantic salmon YOY, similar to the niche complementarity hypothesis, where
388 abundant species have different niches (Mason et al. 2008). In our study, most species had
389 relatively low abundance and low niche overlap or few, weak interactions, thereby partitioning
390 resources and having no impact on Atlantic salmon (Figure 1 – bottom, left quadrant). Having a
391 highly diverse prey community may result in such low levels of competition among species-
392 lifestages (Sánchez-Hernández et al. 2017). A few species-lifestages, like brown trout YOY, had
393 high trophic niche overlap yet low relative abundance and thus strong, but limited interactions
394 and thereby minimal impact to Atlantic salmon (Figure 1 – top, left quadrant). Isotopic niche
395 partitioning has been seen in other systems between fish from the same trophic guild (Dromard
396 et al. 2015; Britton et al. 2018), however, inclusion of relative abundance may show that niche

397 separation is greater between abundant species, as suggested with niche complementarity (Mason
398 et al. 2008). Our methods and conceptual model for combining abundance and trophic niches can
399 be applied in various circumstances in which an ecosystem exhibits a disturbance or addition of a
400 new species (reintroduced or invasive) and could be applied towards other conservation related
401 issues. More studies should incorporate abundance estimates into trophic studies to better
402 understand the impact of interspecific relationships, as we showed here.

403 All three stream communities had similar community isotopic structure and species-
404 lifestage niche overlaps with Atlantic salmon YOY, regardless of having greater or fewer
405 potential competitors. Although there were more instances of co-occurring species-lifestages
406 having high niche overlap with Atlantic salmon YOY in the streams with moderate and high
407 potential competition (Credit River and Cobourg Brook, respectively) than the stream with no
408 potential competitors (Duffins Creek), the overall mean niche overlap of the streams were similar
409 across streams and seasons. From a fish community perspective, the isotopic structure was quite
410 similar across all streams and seasons (excluding post-stocking spring). Thus, there was
411 consistent partitioning of resources among invertivorous fish in the community across seasons
412 and streams, even though the fish communities were different, implying some resource limitation
413 was also occurring. Similar species occurring across streams (e.g., blacknose dace), may have
414 different trophic niche overlaps with Atlantic salmon YOY, yet species partitioned food
415 resources to minimize competition by having either low abundance and high niche overlap
416 (Figure 1 – top, left quadrant) or high abundance and low niche overlap (Figure 1 – bottom, right
417 quadrant). From a community perspective, each stream showed niche complementarity and
418 responded similarly to the addition of Atlantic salmon YOY, regardless of the individual species
419 present. Determining trophic niches and abundance relationships among species may help

420 determine how species organize themselves within communities regarding resource partitioning
421 of the more abundant species and changes in niche space, as well as ecosystem functions and
422 resilience to change.

423 There were some limitations to this study largely related to sampling logistics. With
424 assessments restricted to the post-stocking period, albeit repeatedly over the year, the community
425 may have already shifted to reach a new equilibrium with the reintroduced species, and any
426 response to strong competition (e.g., species with both high overlap and high abundance) had
427 already dissipated. Future research should include a pre- and post-stocking assessment to better
428 quantify how community structure changes or reorganizes following species introduction.
429 Further, due to previous stocking in the streams, the fish community may have already shifted to
430 a new equilibrium prior to this study. Assessing the competitive impact on streams which are
431 stocked with Atlantic salmon YOY for the first time may provide insight on how communities
432 shift with novel disturbances and the duration it may take to reach a new steady state. Using
433 stomach content analyses could also determine immediate community changes with species
434 disruptions, and whether there are stronger competitive impacts upon the disturbance but with
435 lethal sampling could not determine changes over time. Improving upon the methodology of our
436 study will further inform us on the ecological structure and resilience of communities with
437 species reintroductions, among other ecological perturbations.

438 By sampling more streams, we could have assessed a more continuous gradient of
439 community compositions, including an unstocked stream without Atlantic salmon. Although we
440 tried to assess streams with different resident fish communities, and across seasons, some could
441 argue only three sites to be a limitation. All three streams had similar species-lifestage niche
442 overlaps with Atlantic salmon YOY and community isotopic structure, which indicates

consistency across streams in response to a reintroduction. These streams also have similar habitat characteristics typical of those stocked with Atlantic salmon YOY (e.g., cold-water streams with gravel, cobble-boulder streambeds) across Lake Ontario (McKenna and Johnson 2005; Coghlan et al. 2007; Johnson 2013b). Thus, the streams assessed represented conditions Atlantic salmon YOY encounter across Lake Ontario. Future studies could try to incorporate more streams to determine if there are any community composition or habitat-related trends related to niche overlaps with Atlantic salmon and competitive impacts, as well as increase statistical power for different analyses. More importantly, this methodology of determining competitive impact was successfully applied to three different streams and will be beneficial towards assessing community changes with a new species or other disturbances.

In conclusion, we used Lake Ontario Atlantic salmon as an example to show how the inclusion of abundance estimates with trophic niche interactions can be used to assess the competitive impact of the resident community on a reintroduced species. Conventional stable isotope analyses alone may have concluded that brown trout YOY have a high impact on Atlantic salmon YOY, however, the impact was low when abundance was considered. As fish stocking will remain an important management tool, understanding the community impact on the reintroduced species is essential for managing reintroduction success. Our analysis suggests current practices of stocking Atlantic salmon YOY at higher density than resident salmonids is minimising potential competitive impact and thus not limiting the reintroduction of the species. Incorporating abundance improves our understanding of trophic interactions and potentially ecosystem functioning within the community. The stream fish communities assessed supported niche complementarity with low niche overlap among abundant species and Atlantic salmon YOY. Thus, our framework could be applied to other circumstances beyond species

reintroductions such as to better understand potential competitive impacts on other species of interest, and determine whether these impacts influence community structure. It is important to understand species' interspecific interactions through potential competition to determine any biotic limitations that could affect community structure and resource partitioning, particularly with ecological disturbances and species additions.

Acknowledgements

We would like to acknowledge M. Desjardin (OMNRF), J. Ruegg (OMNRF), C. Robinson (OFAH) and Credit Valley Conservation Authority (J. Clayton and C. Bird) for providing information on fish stocking and stocking sites. We thank Ontario Provincial Parks, Durham Board of Education, and Toronto and Region Conservation Authority for granting land access to conduct our study, and Codrington Fisheries Research Facility and W. Sloan (OMNRF) for providing accommodations. We would like to thank E. Halfyard for providing advice on study design. Thank you for all the help in the field and lab: M. Charrone, A. Despenic, T. Fendler, C. Hueval, K. Hudelson, S. Ivanova, K. Johnson, N. Klinard, M. Lucas, S. Pupovac, and A. Weinz. We would like to thank two anonymous reviewers with their constructive comments which greatly improved this paper. All procedures were approved with the University of Windsor Animal Care Committee (permit No. ACC-16-08). Funding: This work was supported by the Natural Sciences and Engineering Research Council of Canada (NSERC) via a Strategic Partnership Grant and Canada Graduate Scholarship.

References

Armstrong, J.D., Kemp, P.S., Kennedy, G.J.A., Ladle, M., and Milner, N.J. 2003. Habitat requirements of Atlantic salmon and brown trout in rivers and streams. *Fish. Res.* **62**: 143–170.

- 489 Britton, J.R., Ruiz-Navarro, A., Verreycken, H., and Amat-Trigo, F. 2018. Trophic consequences
490 of introduced species: Comparative impacts of increased interspecific versus intraspecific
491 competitive interactions. *Funct. Ecol.* **32**(2): 486–495.
- 492 Brown, C., and Day, R. 2002. The future of stock enhancements: lessons for hatchery practice
493 from conservation biology. *Fish Fish.* **3**: 79–94.
- 494 Chen, G., Wu, Z., Gu, B., Liu, D., Li, X., and Wang, Y. 2011. Isotopic niche overlap of two
495 planktivorous fish in southern China. *Limnology* **12**: 151–155.
496 <https://doi.org/10.1007/s10201-010-0332-2>
- 497 Christie, W.J. 1974. Changes in the fish species composition of the Great Lakes. *Journal of the*
498 *Fisheries Research Board of Canada* **31**: 827-854.
- 499 Cochran-Biederman, J.L., Wyman, K.E., French, W.E. and Loppnow, G.L. 2015. Identifying
500 correlates of success and failure of native freshwater fish reintroductions. *Conserv. Biol.*
501 **29**: 175–186.
- 502 Coghlan, S.M. Jr., Cain, G.R., and Ringler, N.H. 2007a. Prey Selection of Subyearling Atlantic
503 Salmon and Rainbow Trout Coexisting in a Natural Stream. *J. Freshw. Ecol.* **22**: 591-607.
- 504 Coghlan, S.M., Connerton, M.J., Ringler, N.H., Stewart, D.J., and Mead, J.V. 2007b. Survival
505 and Growth Responses of Juvenile Salmonines Stocked in Eastern Lake Ontario
506 Tributaries. *Trans. Am. Fish. Soc.* **136**(1): 56–71.
- 507 Crawford, S.S. 2001. Salmonine Introductions to the Laurentian Great Lakes: An Historical
508 Review and Evaluation of Ecological Effects. *Canada Special Publication of Fisheries*
509 *and Aquatic Sciences* 132. NRC Research Press. Ottawa, Ontario. 205 p.

- 510 Dineen, G., Harrison, S.S.C., and Giller, P.S. 2007. Diet partitioning in sympatric Atlantic
511 salmon and brown trout in streams with contrasting riparian vegetation. *J. Fish Biol.*
512 **71**(1): 17–38.
- 513 Dromard, C.R., Bouchon-Navaro, Y., Harmelin-Vivien, M., and Bouchon, C. 2015. Diversity of
514 trophic niches among herbivorous fishes on a Caribbean reef (Guadeloupe, Lesser
515 Antilles), evidenced by stable isotope and gut content analyses. *J. Sea Res.* **95**: 124–131.
- 516 Ewen, J.G., and Armstrong, D.P. 2007. Strategic monitoring of reintroductions in ecological
517 restoration programmes. *Ecoscience* **14**(4): 401–409.
- 518 Fierro, P., Bertrán, C., Mercado, M., Peña-Cortés, F., Tapia, J., Hauenstein, E., Caputo, L.,
519 Vargas-Chacoff, L. 2015. Landscape composition as a determinant of diversity and
520 functional feeding groups of aquatic macroinvertebrates in southern rivers of the
521 Araucanía, Chile. *Latin Am. J. Aquat. Res.* **43**(1): 186–200.
- 522 Gabler, H.M., and Amundsen, P.A. 1999. Resource partitioning between Siberian sculpin
523 (*Cottus poecilopus* Heckel) and Atlantic salmon parr (*Salmo salar* L.) in a sub-Arctic
524 river, northern Norway. *Ecol. Freshw. Fish* **8**(4): 201–208.
- 525 Giller, P.S., and Twomey, H. 1993. Benthic macroinvertebrate community organisation in two
526 contrasting rivers: Between-site differences and seasonal patterns. *Biol. Environ.* **93B**(3):
527 115–125.
- 528 Greig, L., Ritchie, B., Carl, L., and Lewis C.A. 2003. Potential and strategy for restoration of
529 Atlantic Salmon. In. Lake Ontario: A workshop report. Prepared by ESSA Technologies
530 Ltd., Toronto, ON. Ontario Ministry of Natural Resources and Forestry, Lake Ontario
531 Management Unit. Peterborough, pp. 39.

- 532 Griffith, B., Scott, J.M., Carpenter, J.W., and Reed, C. 1989. Translocation as a species
533 conservation tool: status and strategy. *Science* **245**: 477–480.
- 534 Harig, A.L., Fausch, K.D., and Young, M.K. 2000. Factors influencing success of greenback
535 cutthroat trout relocations. *N. Am. J. Fish. Manag.* **20**: 994-1004.
- 536 Hayden, B., Tongnunu, S., Beamish, F.W.H., and Nithirojpakdee, P. 2017. Variation in stable-
537 isotope ratios between fin and muscle tissues can alter assessment of resource use in
538 tropical river. *J. Fish Biol.* **91**: 574–586. doi:10.1111/jfb.13368.
- 539 Heady, W.N., and Moore,
540 J.W. 2013. Tissue turnover and stable isotope clocks to quantify resource shifts in
541 anadromous rainbow trout. *Oecologia* **172**(1): 21–34. doi:10.1007/s00442-012-2483-9.
- 542 Hette-Tronquart, N., Mazeas, L., Reuilly-Manenti, L., Zahm, A., and Belliard, J. 2012. Fish fins
543 as non-lethal surrogates for muscle tissues in freshwater food web studies using stable
544 isotopes. *Rapid Commun. Mass Spectrom.* **26**(14): 1603–1608.
- 545 Houde, A.L.S., Wilson, C.C., and Neff, B.D. 2015a. Competitive interactions among multiple
546 non-native salmonids and two populations of Atlantic salmon, *Ecol. Freshw. Fish* **24**: 44–
547 55.
- 548 Houde, A.L.S., Wilson, C.C., and Neff, B.D. 2015b. Effects of competition with four non-native
549 salmonid species on Atlantic salmon from three populations. *Trans. Am. Fish. Soc.*
550 **144**:1081-1090.
- 551 Houde, A.L.S., Smith, A.D., Wilson, C.C., Peres-Neto, P.R., and Neff, B.D. 2016. Competitive
552 effects between rainbow trout and Atlantic salmon in natural and artificial streams. *Ecol.*
Freshw. Fish **25**(2): 248–260.

- 553 Houde, A.L.S., Wilson, C.C., and Neff, B.D. 2017. Performance of four salmonids species in
554 competition with Atlantic salmon. *J. Great Lakes Res.* **43**: 211–215.
- 555 Hutchinson, G.E. 1957. Concluding remarks. *Cold Spring Harb. Symp. Quant. Biol.* **22**: 415–
556 427.
- 557 Jachowski, D.S., Millsbaugh, J.J., Angermeier, P.L., and Slotow, R. 2016. Reintroduction of Fish
558 and Wildlife Populations. University of California Press.
- 559 Jackson, A.L., Inger, R., Parnell, A.C., and Bearhop, S. 2011. Comparing isotopic niche widths
560 among and within communities: SIBER – Stable Isotope Bayesian Ellipses in R. *J. Anim.*
561 *Ecol.* **80**: 595–602.
- 562 Johnson, J.H. 2013a. Diel variation in summer habitat use, feeding periodicity, and diet of
563 subyearling Atlantic salmon in the Salmon River Basin, New York. *J. Great Lakes Res.*
564 **39**(3): 493–498.
- 565 Johnson, J.H. 2013b. Habitat use and diet composition of juvenile Atlantic salmon in a tributary
566 of Lake Ontario. *J. Great Lakes Res.* **39**: 162 – 167.
- 567 Johnson, J.H., and McKenna, J.E. 2015. Diel Resource Partitioning among Juvenile Atlantic
568 Salmon, Brown Trout, and Rainbow Trout during Summer. *N. Am. J. Fish. Manag.*
569 **35**(3): 586–597.
- 570 Johnson, J.H., DiRado, J.A., Mackey, G., and Abbett, R. 2017. Comparative diets and foraging
571 strategies of subyearling Atlantic salmon, brown trout, and rainbow trout during winter. *J.*
572 *Appl. Ichthyol.* **33**: 1158–1165.

- 573 Kawaguchi, Y., and S. Nakano. 2001. Contribution of terrestrial invertebrates to the annual
574 resource budget for salmonids in forest and grassland reaches of a headwater stream.
575 Freshw. Biol. **46**(3): 303–316.
- 576 Keeley, E.R., and Grant, J.W.A. 1997. Allometry of diet selectivity in juvenile Atlantic salmon
577 (*Salmo salar*). Can. J. Fish. Aquat. Sci. **54**(8): 1894–1902.
- 578 Kiljunen, M., Grey, J., Sinisalo, T., Harrod, C., Immonen, H., and Jones, R.I., 2006. A revised
579 model for lipid-normalizing $\delta^{13}\text{C}$ values from aquatic organisms, with implications for
580 isotope mixing models. J. Appl. Ecol. **43**: 1213–1222.
- 581 Layman, C.A., Arrington, D.A., Montana, C.G., and Post, D.M. 2007. Can stable isotope ratios
582 provide for community-wide measures of trophic structure? Ecology **88**(1): 42–48.
- 583 Levin, P.S., Achord, S., Feist, B.E. and Zabel, R.W. 2002. Non-indigenous brook trout and the
584 demise of Pacific salmon: a forgotten threat? P. Roy. Soc. B. **269**: 1663–1670.
- 585 MacArthur, R. and Levins, R. 1967. The limiting similarity, convergence, and divergence of
586 coexisting species. Amer. Nat. **101**: 377–385.
- 587 Mason, N.W.H., Lanoiselée, C., Mouillot, D., Wilson, J.B., and Argillier, C. 2008. Does niche
588 overlap control relative abundance in French lacustrine fish communities? A new method
589 incorporating functional traits. J. Anim. Ecol. **77**: 661–669.
- 590 McIntyre, P.B., and Flecker, A.S. 2006. Rapid turnover of tissue nitrogen of primary consumers
591 in tropical freshwaters. Oecologia **148**(1): 12–21. doi:10.1007/s00442-005-0354-3.
- 592 McKenna, J.E., and Johnson, J.H. 2005. Juvenile Rainbow Trout Production in New York
593 Tributaries of Lake Ontario: Implications for Atlantic Salmon Restoration. North Am. J.
594 Fish. Manag. **25**(1): 391–403. doi:10.1577/m04-022.1.

- 595 Minckley, W.L. 1995. Translocation as a tool for conserving imperiled fishes: experiences in
596 western United States. *Biol. Conserv.* **72**: 297–309.
- 597 Mittelbach, G.G., and Persson, L. 1998. The ontogeny of piscivory and its ecological
598 consequences. *Can. J. Fish. Aquat. Sci.* **55**(6): 1454–1465.
- 599 Molony, B.W., Lenanton, R., Jackson, G., and Norriss, J. 2003. Stock enhancement as a fisheries
600 management tool. *Rev. Fish Biol. Fish.* **13**: 409–432.
- 601 Nakano, S., H. Miyasaka, and N. Kuhara. 1999. Terrestrial-aquatic linkages: Riparian arthropod
602 inputs alter trophic cascades in a stream food web. *Ecol.* **80**(7): 2435–2441.
- 603 Ogloff, W.R., Yurkowski, D.J., Davoren, G.K., and Ferguson, S.H. 2019. Diet and isotopic niche
604 overlap elucidate competition potential between seasonally sympatric phocids in the
605 Canadian Arctic. *Mar. Biol.* **166**:103.
- 606 Peterson, B.J., and Fry, B. 1987. Stable Isotopes in Ecosystem Studies. *Annu. Rev. Ecol. Evol.*
607 *Syst.* **18**: 293–320.
- 608 Post, D.M., Layman, C.A., Arrington, D.A., Takimoto, G., Quattrochi, J., and Montana, C.G.,
609 2007. Getting to the fat of the matter: models, methods, and assumptions for dealing with
610 lipids in stable isotope analyses. *Oecologia* **152**: 179-189.
- 611 R Core Team. 2019. R: A language and environment for statistical computing. R Foundation for
612 Statistical Computing, Vienna, Austria. URL <https://www.R-project.org/>.
- 613 Robinson, A.T., and Ward, D.L. 2011. Interactions between Desert Pupfish and Gila Topminnow
614 Can Affect Reintroduction Success. *N. Am. J. Fish. Manag.* **31**(6): 1093–1099.

- 615 Sagouis, A., Cucherousset, J., Villéger, S., Santoul, F., and Boulêtreau, S. 2015. Non-native
616 species modify the isotopic structure of freshwater fish communities across the globe.
617 *Ecography* **38**: 979-985.
- 618 Sánchez-Hernández, J., Gabler, H. and Amundsen, P. 2016. Food resource partitioning between
619 stream-dwelling Arctic charr *Salvelinus alpinus* (L.), Atlantic salmon *Salmo salar* L. and
620 alpine bullhead *Cottus poecilopus* Heckel, 1836: an example of water column
621 segregation. *Hydrobiologia* **783**: 105–115.
- 622 Sánchez-Hernández, J., Gabler, H.M., and Amundsen, P.A. 2017. Prey diversity as a driver of
623 resource partitioning between river-dwelling fish species. *Ecol. Evol.* **7**(7): 2058–2068.
- 624 Sanderson, B.L., Tran, C.D., Coe, H.J., Pelekis, V., Steel, E.A., and Reichert, W.L. 2009.
625 Nonlethal sampling of fish caudal fins yields valuable stable isotope data for threatened
626 and endangered fishes. *Trans. Am. Fish. Soc.* **138**(5): 1166–1177.
- 627 Schooley, J.D., and Marsh, P.C. 2007. Stocking of endangered razorback suckers in the lower
628 Colorado River basin over three decades: 1974–2004. *N. Am. J. Fish. Manag.* **27**: 43–51.
- 629 Scott, W.B., and Crossman, E.J. 1998. Freshwater fishes of Canada. Galt House Publications,
630 Oakville, Canada.
- 631 Scott, R.J., Noakes, D.L.G., Beamish, F.W.H., and Carl, L.M., 2003. Chinook salmon impede
632 Atlantic salmon conservation in Lake Ontario. *Ecol. Freshw. Fish* **12**: 66–73.
- 633 Shustov, Y.A., Tyrkin, I.A., and Rasputina, Y.N. 2014. Trophic Competition of Common
634 Sculpin and Stone Loach with Juvenile Salmon of Gen. *Salmo* in Tributaries of the
635 Onega Lake. *Hydrobiol. J.* **50**:32-40.

- 636 Skinner, M.M., Martin, A.A., and Moore, B.C., 2016. Is lipid correction necessary in the stable
637 isotope analysis of fish tissues? *Rapid Commun. Mass Spectrom.* **30**: 881-889.
- 638 Stanfield, L., and Jones, M.L. 2003. Factors influencing rearing success of Atlantic salmon
639 stocked as fry and parr in Lake Ontario tributaries. *N. Am. J. Fish. Manag.* **23**: 1175 –
640 1183.
- 641 Sugihara, G., Bersier, L.F., Southwood, T.R.E., Pimm, S.L. and May, R.M. 2003. Predicted
642 correspondence between species abundances and dendro- grams of niche similarities.
643 *PNAS USA* **100**: 5246–5251.
- 644 Van Zwol, J.A., Neff, B.D., and Wilson, C.C. 2012. The effect of competition among three
645 salmonids on dominance and growth during the juvenile life stage. *Ecol. Freshw.* **21**(4):
646 533–540.
- 647 Wang, J., Chapman, D., Xu, J., Wang, Y., and Gu, B. 2018. Isotope niche dimension and trophic
648 overlap between bigheaded carps and native filter-feeding fish in the lower Missouri
649 River, USA. *PLOS ONE* **13**(6): e0199805.
- 650 Ward, D.M., Nislow, K.H., and Folt, C.L. 2008. Do native species limit survival of reintroduced
651 Atlantic salmon in historic rearing streams? *Biol. Conserv.* **141**(1): 146–152.

Can. J. Fish. Aquat. Sci. Downloaded from cdnsciencepub.com by UNIV WINDSOR on 03/08/21
For personal use only. This Just-IN manuscript is the accepted manuscript prior to copy editing and page composition. It may differ from the final official version of record.

Tables

Table 1. The number of times a species-lifestage was captured simultaneously (or co-occurring) with Atlantic salmon young-of-year (YOY) across seasons (excluding spring) and the number and proportion of the 40% isotopic niche overlaps that were greater than 0.5, grouped by location and species-lifestages. Trophic niche overlaps refers to either Atlantic salmon YOY overlapping a species-lifestage or vice versa and was counted once per co-occurrence (e.g., if both methods of overlap were > 0.5 it was only counted once).

Location	Co-occurrences	Overlaps > 0.5	Proportion of overlaps > 0.5
Cobourg Brook	16	5	0.31
Credit River	16	5	0.31
Duffins Creek	13	2	0.15

Species-lifestages	Co-occurrences	Overlaps > 0.5	Proportion of overlaps > 0.5
Atlantic Salmon – Yearling	8	2	0.25
Brook Trout – YOY	1	0	0.00
Brook Trout – Adult	1	0	0.00
Brown Trout – YOY	4	3	0.75
Brown Trout – Yearling	2	1	0.50
Rainbow Trout – YOY	3	1	0.33
Rainbow Trout – Yearling	3	0	0.00
Blacknose Dace – Adult	6	3	0.50
Longnose Dace – Adult	5	0	0.00
Rainbow Darter – Adult	5	2	0.40
Mottled Sculpin – Adult	3	0	0.00
Slimy Sculpin – Adult	3	0	0.00
White Sucker – Adult	1	0	0.00

Figure captions

Figure 1. Conceptual interspecific interaction-abundance impact model based on trophic niche overlap between species X and other species in the community, and the relative abundance of other species in the community. Grey scale indicates potential level of impact to species X based on interaction strength of the niche overlap and number of potential interactions from other species' relative abundance if greater or lesser than relative abundance of species X, where light grey indicates negligible impact, grey is low to moderate impact, and dark grey is high impact.

Figure 2. The A) total and species-lifestage catch-per-unit-effort (CPUE; number captured per electrofishing seconds) and B) relative abundance of species-lifestage in each stream (Co = Cobourg Brook; Cr = Credit River; Du = Duffins Creek) and season (Sp = spring; Su = summer; Fa = fall; Ow = overwinter). Stream barplots are segregated by black lines.

Figure 3. Stable isotope bi-plot of the isotopic niches of species-lifestages of invertivores in the fish community by season and location. Thick circles enclose standard (40%) ellipse areas for each species-lifestages. Hatchery Atlantic Salmon signatures are portrayed in spring for each location for reference.

Figure 4. Mean \pm SE of the proportion overlap of the standard ellipse area with A) Atlantic salmon young-of-year (ATLS) overlapping species-lifestages (SL) and B) SL overlapping ATLS, by stream and season.

Figure 5. The proportion overlap of the standard ellipse area with A) Atlantic salmon young-of-year (ATLS) overlapping species-lifestages (SL) and B) SL overlapping ATLS by the relative abundance of SL. Grey scale indicates the potential level of impact to ATLS based on interaction strength of the niche overlap as depicted in Figure 1. High niche overlap is considered > 0.5 , and the lowest relative abundance of ATLS across sampling events was 0.12.

Figure 6. Density plots showing the isotopic structure or Bayesian Layman's metrics (TA = total convex hull area; CD = mean distance from the centroid; NND = mean nearest neighbour distance, SDNND = standard deviation of the nearest neighbour distance) for the fish community in each stream (Co = Cobourg Brook, Cr = Credit River, Du = Duffins Creek) and season (Sp = spring, Su = summer, Fa = fall, Ow = overwinter). The black dots represent the mode and boxed areas reflect the 95, 75 and 50% credible intervals.

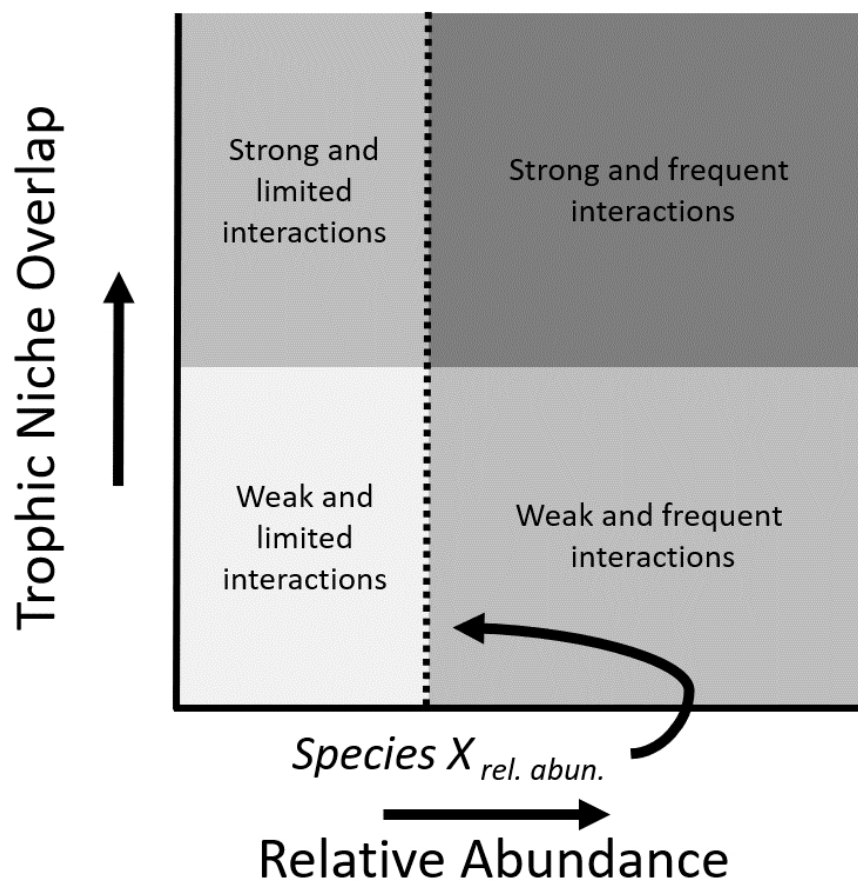


Figure 1. Conceptual interspecific interaction-abundance impact model based on trophic niche overlap between species X and other species in the community, and the relative abundance of other species in the community. Grey scale indicates potential level of impact to species X based on interaction strength of the niche overlap and number of potential interactions from other species' relative abundance if greater or lesser than relative abundance of species X, where light grey indicates negligible impact, grey is low to moderate impact, and dark grey is high impact.

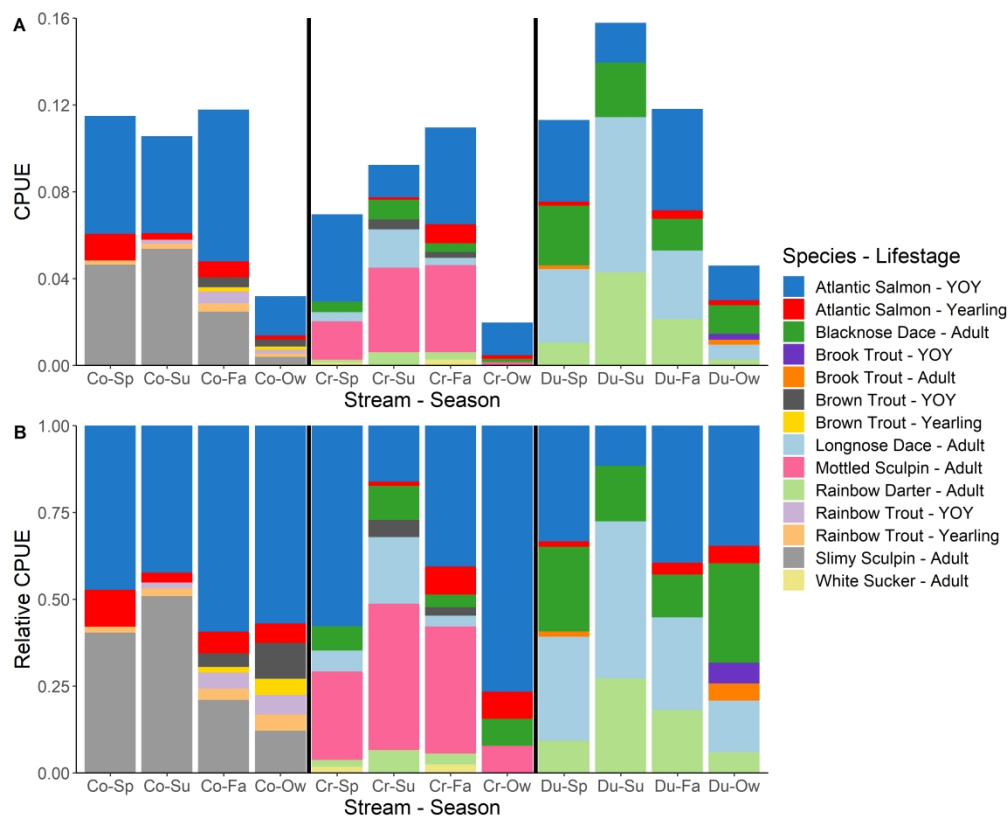


Figure 2. The A) total and species-lifestage catch-per-unit-effort (CPUE; number captured per electrofishing seconds) and B) relative abundance of species-lifestage in each stream (Co = Cobourg Brook; Cr = Credit River; Du = Duffins Creek) and season (Sp = spring; Su = summer; Fa = fall; Ow = overwinter). Stream barplots are segregated by black lines.

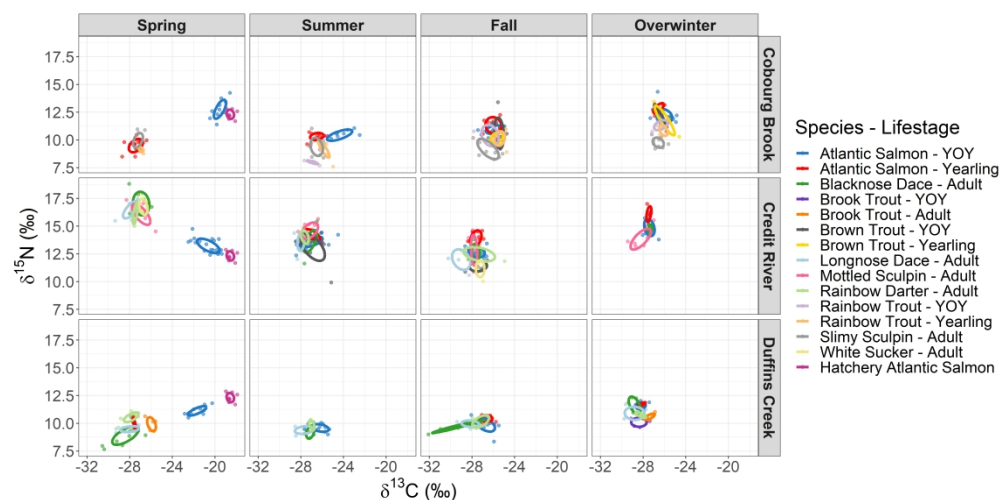


Figure 3. Stable isotope bi-plot of the isotopic niches of species-lifestages of invertivores in the fish community by season and location. Thick circles enclose standard (40%) ellipse areas for each species-lifestages. Hatchery Atlantic Salmon signatures are portrayed in spring for each location for reference.

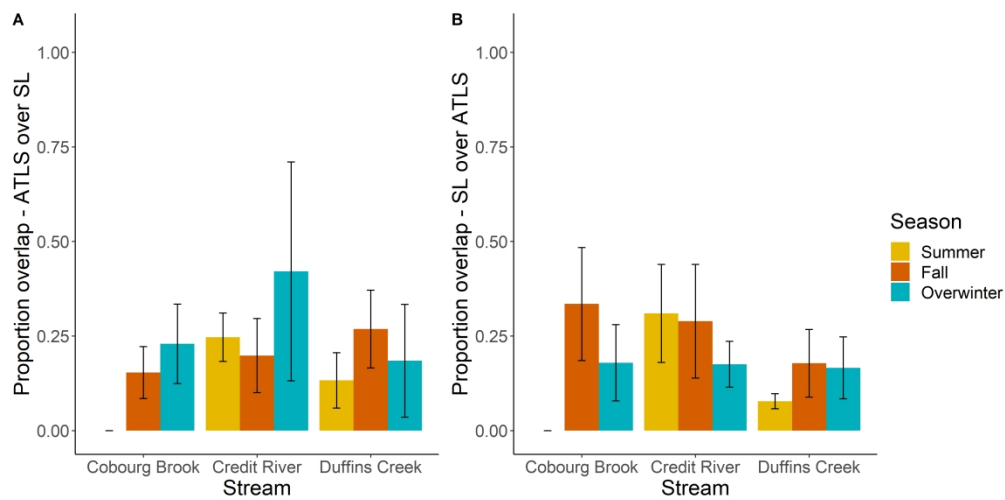


Figure 4. Mean \pm SE of the proportion overlap of the standard ellipse area with A) Atlantic salmon young-of-year (ATLS) overlapping species-lifestages (SL) and B) SL overlapping ATLs, by stream and season.

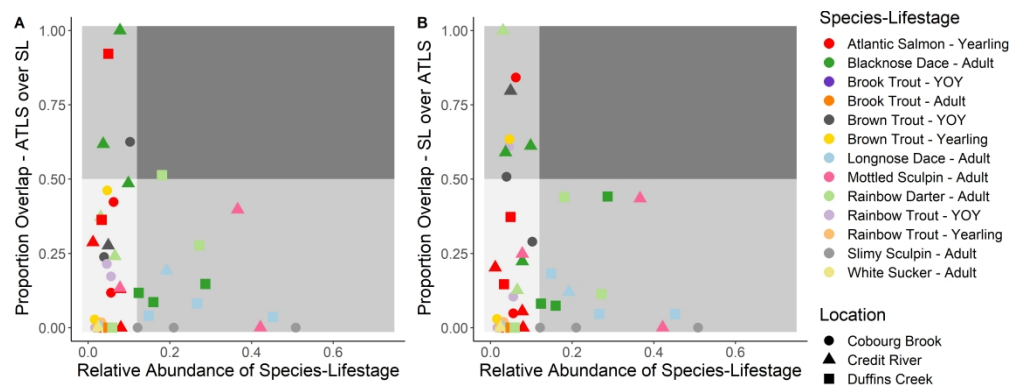


Figure 5. The proportion overlap of the standard ellipse area with A) Atlantic salmon young-of-year (ATLS) overlapping species-lifestages (SL) and B) SL overlapping ATLS by the relative abundance of SL. Grey scale indicates the potential level of impact to ATLS based on interaction strength of the niche overlap as depicted in Figure 1. High niche overlap is considered > 0.5 , and the lowest relative abundance of ATLS across sampling events was 0.12.

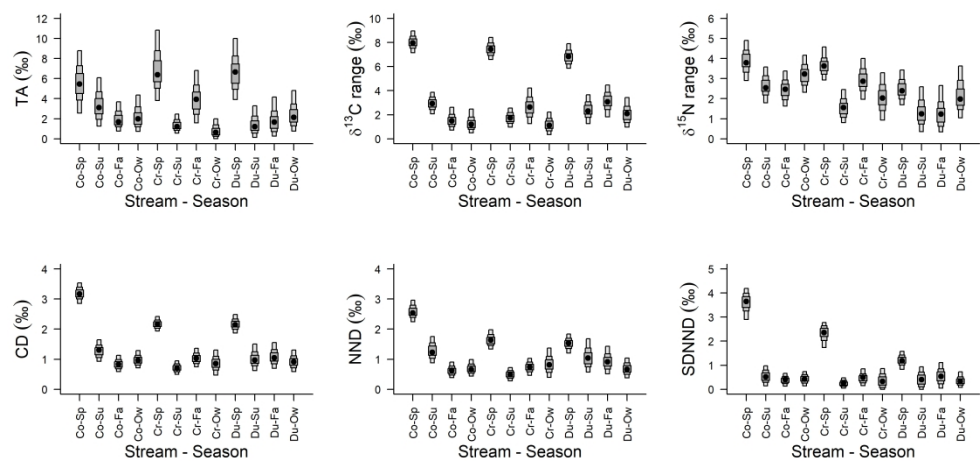


Figure 6. Density plots showing the isotopic structure or Bayesian Layman’s metrics (TA = total convex hull area; CD = mean distance from the centroid; NND = mean nearest neighbour distance, SDNND = standard deviation of the nearest neighbour distance) for the fish community in each stream (Co = Cobourg Brook, Cr = Credit River, Du = Duffins Creek) and season (Sp = spring, Su = summer, Fa = fall, Ow = overwinter). The black dots represent the mode and boxed areas reflect the 95, 75 and 50% credible intervals.